

第四届全国稳定同位素生态学学术研讨会

我国北方城市冬季PM_{2.5}暴增机制 及相关液相反应

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2017年10月16-18日 南京信息工程大学

提 纲

1. 冬季雾霾期硫酸盐的快速增长
 - (1) 北京西安外场观测
 - (2) 实验室烟雾箱模拟
2. 沙尘暴期二次粒子
 - (1) 硝酸铵
 - (2) 二次有机气溶胶
3. 区域生物质燃烧
 - (1) 无机盐
 - (2) 有机酸

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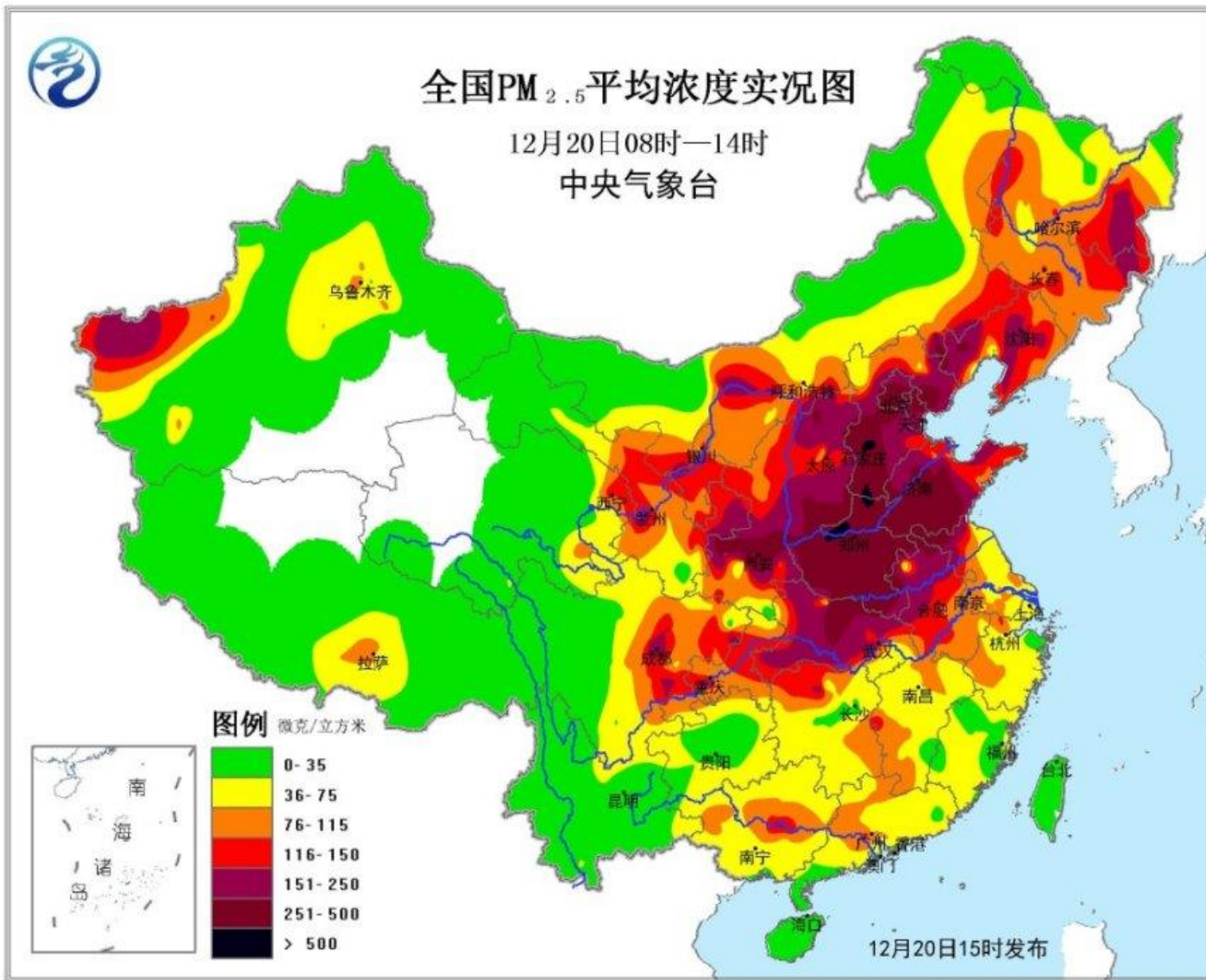
(2) 二次有机气溶胶

3. 区域生物质燃烧

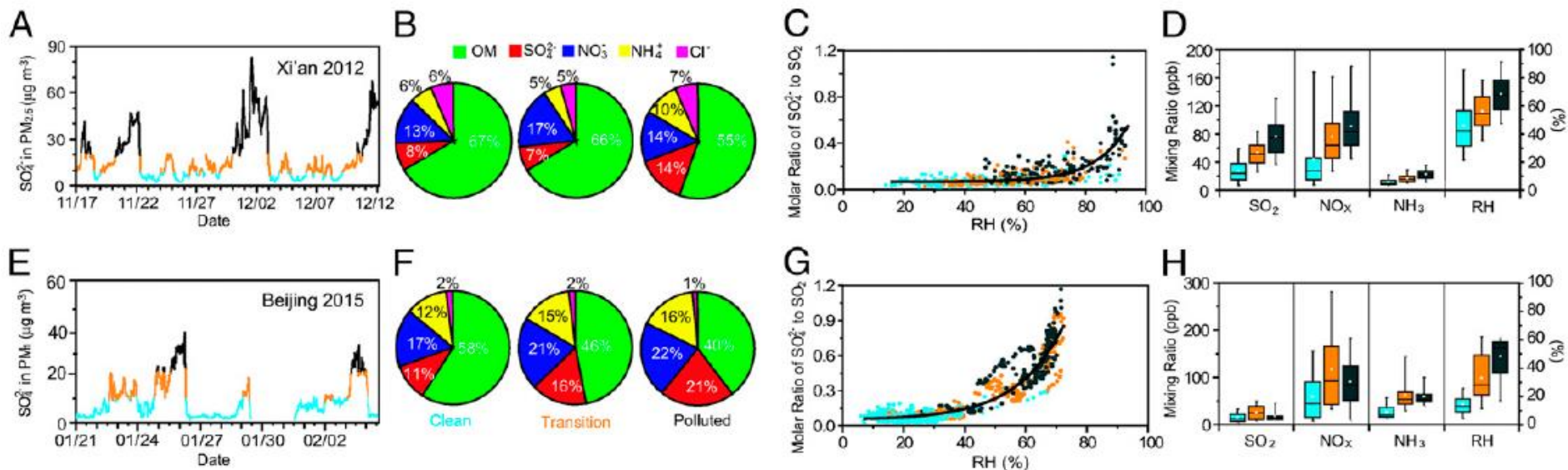
(1) 无机盐

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1. 冬季雾霾期硫酸盐的快速增长



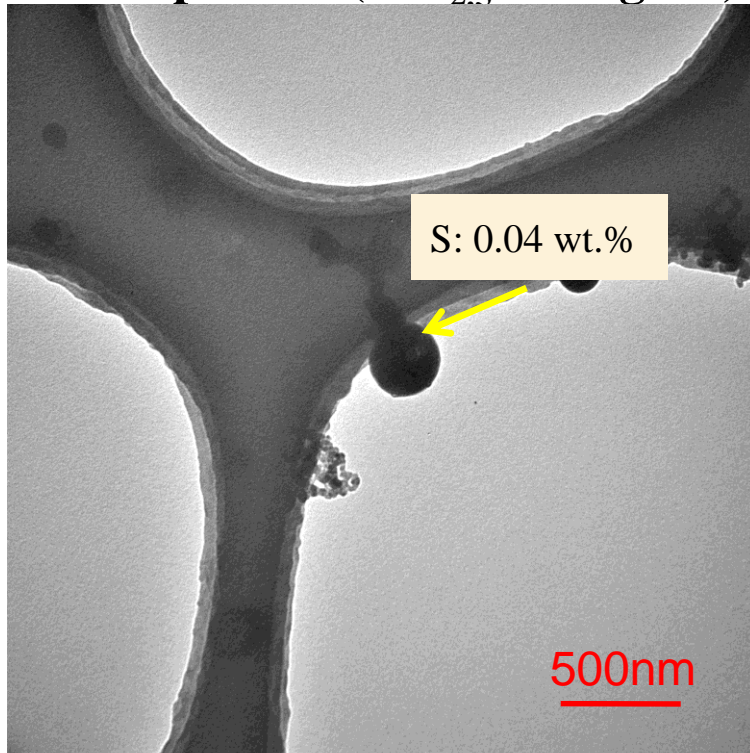
西安和北京冬季灰霾期间硫酸盐形成与演化 (Wang et al., PNAS 2016)



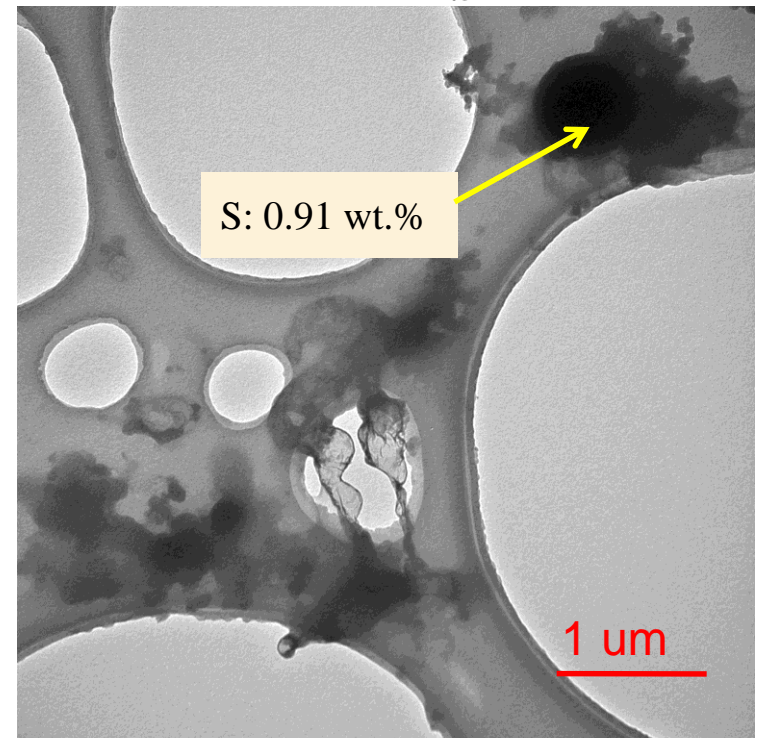
- (1) 清洁期：有机物为主，污染期：无机盐为主
- (2) 西安有机物含量高于北京
- (3) 硫酸盐随着相对湿度的增加而呈指数型增长
- (4) 北京 SO_2 可在更低相对湿度下转化成硫酸盐
- (5) 重污染期间伴随着高浓度 SO_2 , NO_x , NH_3 以及高相对湿度

扫描电镜分析

Clean period ($\text{PM}_{2.5} < 75 \text{ ug/m}^3$)



Hazy period ($\text{PM}_{2.5} > 200 \text{ ug/m}^3$)

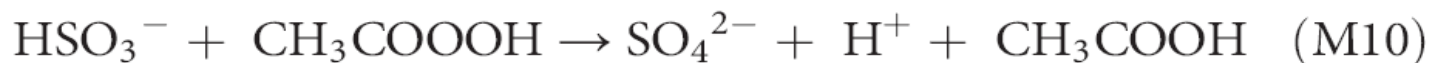
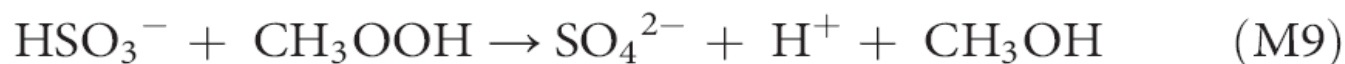
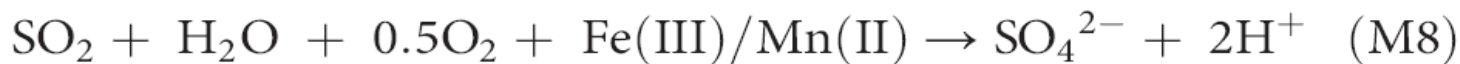
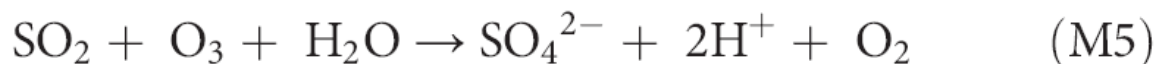
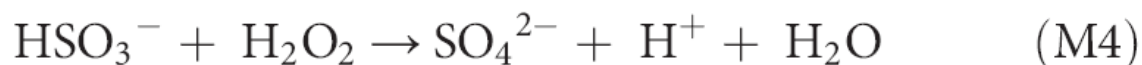


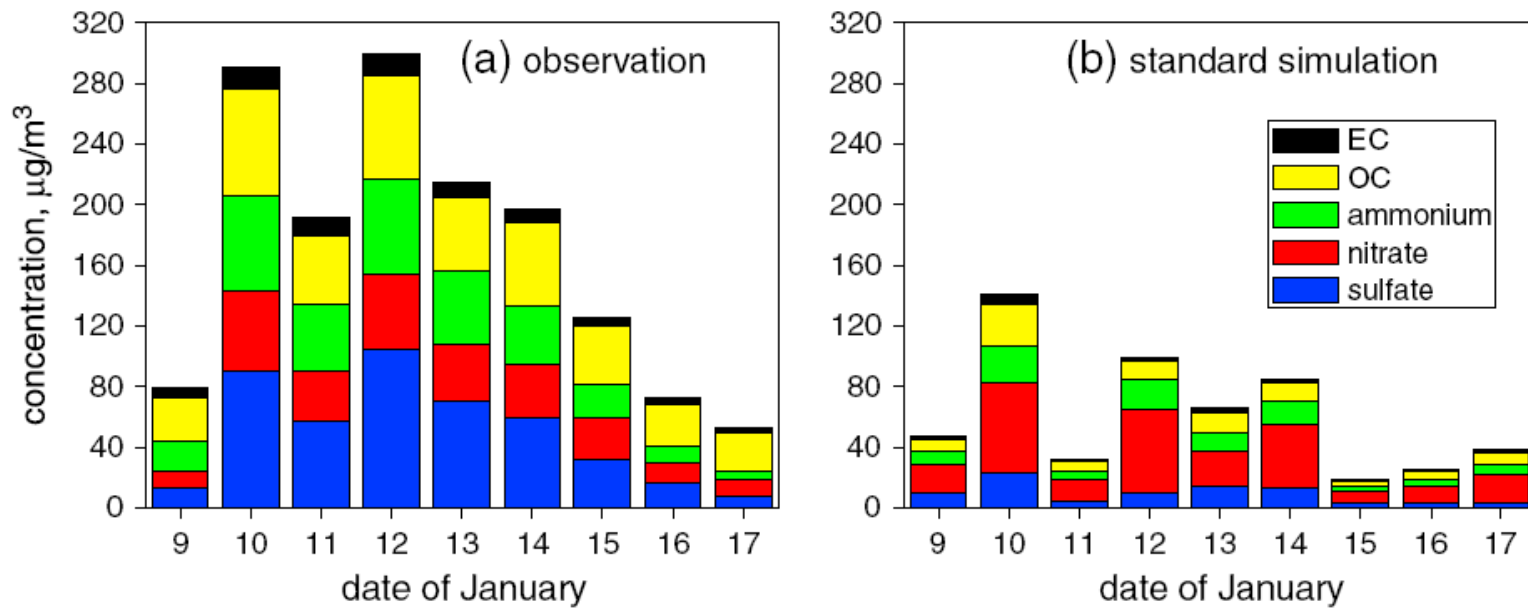
- (1) The content of sulfur in particles, which is SO_4^{2-} , is around 20 times higher on hazy days in comparison with that in clean periods;
- (2) Most of the aerosols on hazy days are liquid

Again indicating the aerosol aqueous phase formation of sulfate in haze episodes



液相反应机制



**Table 2.** A Summary of All the Simulations in This Study

Simulation Name	Description
Standard simulation	Model standard simulation with the 2013 inventory
DoubleSO ₂ run	SO ₂ emissions from the standard simulation are doubled over NC and reduced by 30% over SC. Changes are applied uniformly throughout January.
Emission_run	Total monthly emissions of SO ₂ , OC, and EC (anthropogenic portions only) from the standard run are increased by 100% over NC; SO ₂ emissions over SC are reduced by 30%; NO _x and NH ₃ emissions are the same as those in the standard run. The meteorology correction factors are applied on a day-to-day basis to emissions of SO ₂ , NO _x , NH ₃ , OC, and EC over NC.
Gamma_run T1	Emissions are the same as in the emission_run; $\gamma = 10^{-4}$ when RH = 50%, $\gamma = 10^{-3}$ when RH = 50%; γ increases linearly with RH from 50% to 100% (equation (2)).
Gamma_run T2	Emissions are the same as those in the emission_run; $\gamma = 10^{-3}$ when RH = 50%, $\gamma = 10^{-2}$ when RH = 50%; γ increases linearly with RH from 50% to 100% (equation (2)).
Gamma_run T3	Emissions are the same as those in the emission_run; $\gamma = 10^{-2}$ when RH = 50%, $\gamma = 10^{-1}$ when RH = 50%; γ increases linearly with RH from 50% to 100% (equation (2)).

西安2012年冬季高时间分辨率滤膜采样分析显示： 过渡金属铁锰氧化不重要

Table S1. Gaseous and PM pollutants and meteorological parameters during Xi'an 2012

	Clean		Transition		Polluted	
	Mean	Range	Mean	Range	Mean	Range
I. Gaseous pollutants (ppb)						
SO ₂	28±17	1.0-86	54±22	17-191	78±31	16-203
NO _x	44±49	5.0-264	76±51	15-300	92±39	25-245
O ₃	7.4±7.0	0.0-26	4.1±4.7	0.4-11	4.1±2.4	0.6-9.6
NH ₃	12±7.4	4.7-67	17±7.7	7.6-35	23±8.3	9.3-61
HONO	1.3±1.0	0.2-5.4	2.1±1.3	0.2-6.5	2.7±1.8	0.3-10
II. Inorganic ions, Fe, Mn and organic matter in PM_{2.5} (μg m⁻³)						
SO ₄ ²⁻	5.9±2.2	2.3-10	14±4.4	10-20	38±14	20-83
NO ₃ ⁻	8.7±4.9	1.4-25	16±6.7	3.8-35	33±10	12-55
Cl ⁻	4.0±3.7	0.0-22	9.8±5.1	2.4-28	14±6.3	2.6-34
NH ₄ ⁺	4.0±2.2	0.8-11	10±3.7	5.1-18	25±7.7	3.2-44
Na ⁺	3.6±3.2	0.2-8.4	4.5±3.2	0.5-17	4.2±2.7	0.5-17
K ⁺	1.3±0.7	0.3-4.1	3.1±1.2	1.3-7.0	4.6±1.4	1.8-8.3
Mg ²⁺	0.2±0.1	0.1-0.7	0.3±0.1	0.0-0.7	0.3±0.1	0.0-0.8
Ca ²⁺	1.6±1.0	0.3-6.3	2.4±1.2	0.0-5.3	2.3±1.2	0.2-5.9
Total ions	20±13	6.8-63	60±10	34-97	121±32	65-199
Fe (μg m ⁻³)	0.82±0.29	0.37-1.13	1.51±0.70	0.60-3.0	1.76±0.66	0.79-2.79
Mn (μg m ⁻³)	0.04±0.04	0.00-0.10	0.11±0.08	0.04-0.35	0.15±0.07	0.08-0.29
Water-soluble Fe (ng m ⁻³)	1.5±2.1	0.0-6.1	4.6±3.9	0.0-14	16±5.1	7.3-23
Water-soluble Mn (ng m ⁻³)	10±2.1	3.8-20	21±8.7	11-40	41±16	17-70
Organic matter (OM)	35±15	7.0-70	99±33	38-163	177±39	116-288
pH	6.70±1.40	4.43-11.0	6.04±1.24	4.16-8.03	6.96±1.33	4.14-8.16
III. PM_{2.5} and meteorological parameters						
PM _{2.5} (μg m ⁻³)	43±18	8.0-74	139±65	76-613	250±120	101-839
T (°C)	5.7±4.1	-2.0-17	4.1±4.0	-2.3-11	4.1±4.4	-3.1-14
RH (%)	46±18	14-94	56±17	26-93	68±14	41-93
Visibility (km)	8.9±3.4	3.2-17	6.1±2.8	2.4-12	3.2±1.1	1.4-7.2

实验室反应瓶模拟：大气中O₂没有效应

Table S3. Detection of sulfate formation in the reaction cell

Experimental run	SO ₂ (350 ppm)	NO ₂ (350 ppm)	Water	3 wt % NH ₃	Integrated sulfate desorption peak area (x 10 ⁶ cps)
1 (3)	In N ₂	In N ₂	√	x	6.8±2.6
2 (3)	In N ₂	In N ₂	x	√	11.0±4.3
3 (1)	In air	In air	√	x	6.5
4 (1)	In air	In air	x	√	10.0

The symbols “√” and “x” indicate whether a water or NH₃ solution was used and not used in the exposure, respectively. The number in parenthesis on the right column denotes the number of repeating experiments.

确认：无光照条件下：

(1)二氧化氮可以快速液相氧化二氧化硫，生成硫酸盐

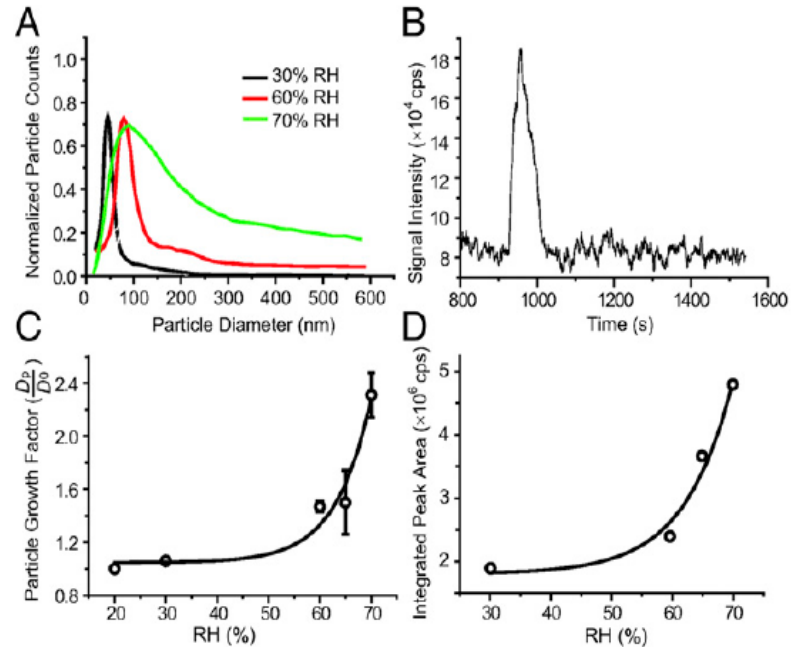
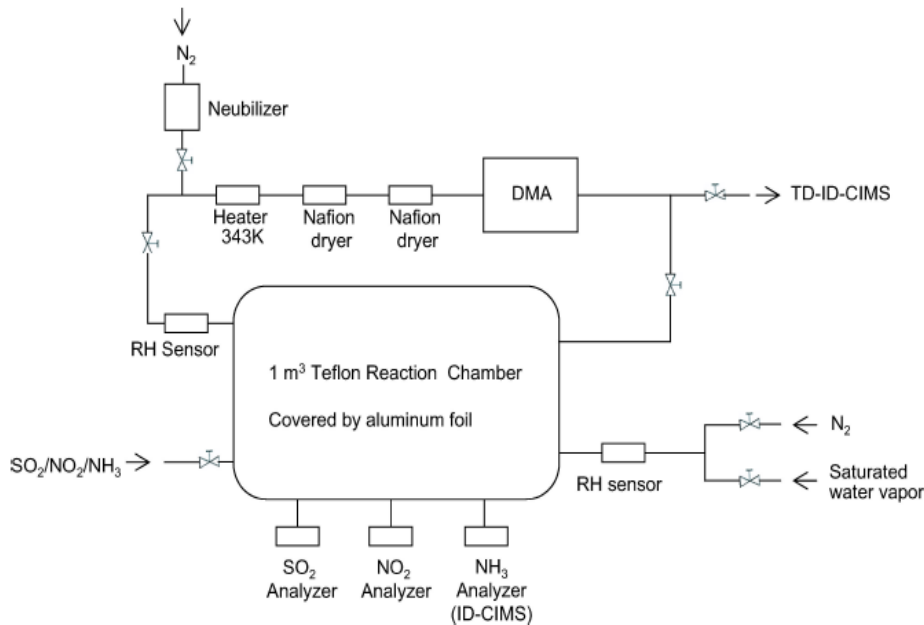
(2)氨气可促进上述反应进行

(3)在此非均相反应过程中，大气中氧气没有作用

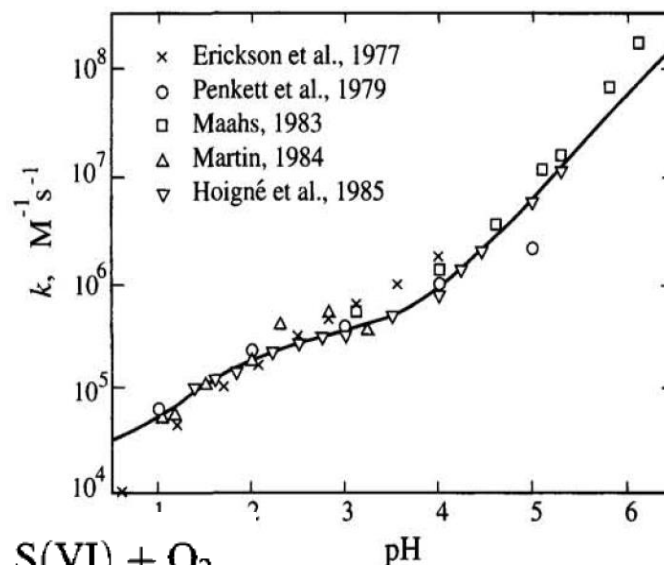
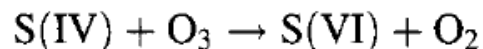
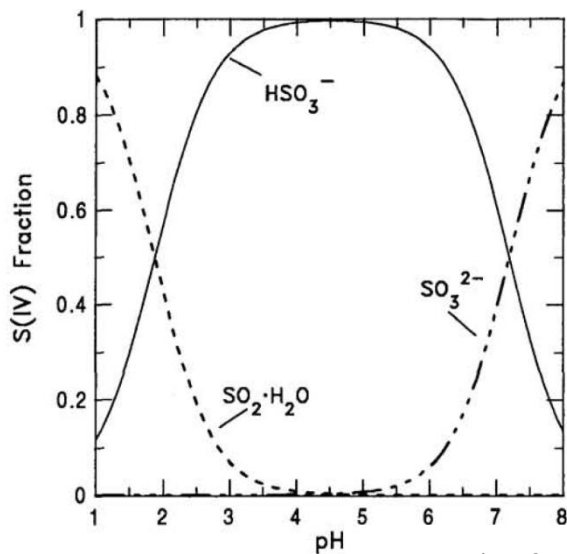
烟雾箱模拟：纳米粒子上液相 $\text{SO}_2 + \text{NO}_2$ 氧化(重要)



1立方米烟雾箱，通入二氧化硫、二氧化氮、氨气和水汽（调节相对湿度），45 nm seed (草酸)
1个小时后，用仪器检测硫酸盐的浓度，发现：
当二氧化硫、二氧化氮、氨气、>60%相对湿度四个条件同时满足时，硫酸盐才能形成，并且与相对湿度成指数增长关系；
没有氨气是看不到硫酸盐的信号，



二氧化硫氧化形成硫酸盐颗粒物酸度是关键，氨气对酸度起着决定性作用

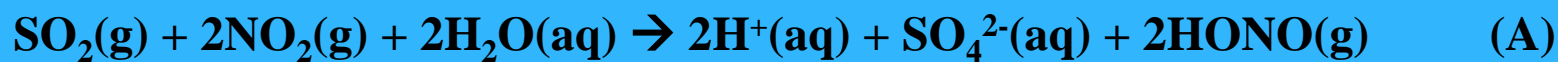
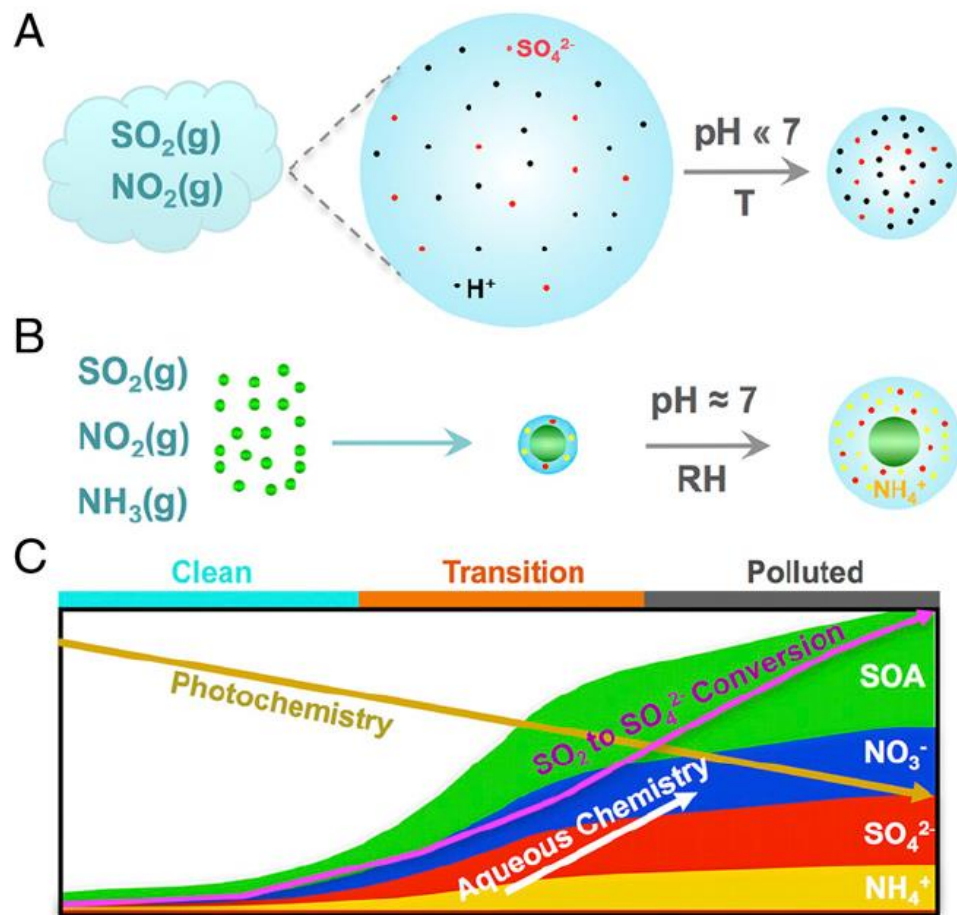


$$R_0 = -\frac{d[\text{S(IV)}]}{dt} = (k_0[\text{SO}_2 \cdot \text{H}_2\text{O}] + k_1[\text{HSO}_3^-] + k_2[\text{SO}_3^{2-}])[\text{O}_3] \quad (7.80)$$

with $k_0 = 2.4 \pm 1.1 \times 10^4 \text{ M}^{-1} \text{ s}^{-1}$, $k_1 = 3.7 \pm 0.7 \times 10^5 \text{ M}^{-1} \text{ s}^{-1}$ and, $k_2 = 1.5 \pm 0.6 \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$. The activation energies recommended by Hoffmann and Calvert (1985) are based on the work of Erickson et al. (1977) and are 46.0 kJ mol^{-1} for k_1 and 43.9 kJ mol^{-1} for k_2 .

(Seinfeld and Pandis., 1998)

Wang et al., PNAS 2016 : 伦敦雾与中国霾

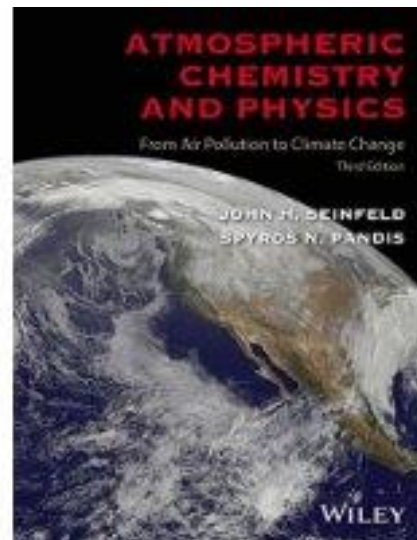


本项工作的重要性

We see that under these conditions oxidation by dissolved H_2O_2 is the predominant pathway for sulfate formation at pH values less than roughly 4–5. At $\text{pH} \geq 5$ oxidation by O_3 starts dominating and at pH 6 it is 10 times faster than that by H_2O_2 . Also, oxidation of S(IV) by O_2 catalyzed by Fe and Mn may be important at high pH, but uncertainties in the rate expressions at high pH preclude a definite conclusion. **Oxidation of S(IV) by NO_2 is unimportant at all pH for the concentration levels above.**

----Seinfeld and Pandis “*Atmospheric Chemistry and Physics*”

tion process with enforced NH_3 and NO_2 control measures. In addition to explaining the polluted episodes currently occurring in China and during the 1952 London Fog, this sulfate production mechanism is widespread, and our results suggest a way to tackle this growing problem in China and much of the developing world.



On Sep 27, 2016, at 12:03 PM, Molina, Mario <mjmolina@ucsd.edu> wrote:

Hi Renyi,

Sorry it has taken so long to respond. It turns out I had a couple of urgent projects to finish here at our study center, and they took longer than anticipated.

Anyhow, I am working on the Sulphur paper, and I agree that it is an important piece of work on atmospheric chemistry. And yet, given the various commentaries by the reviewers I

**M. Molina, 1995年
大气化学诺贝尔奖得主**

灰霾成因论文 (Wang et al., PNAS 2016) 受到李克强总理、 全国政协副主席罗富和等国家领导人高度关注，设立总理 基金

周卫健院士建议攻克雾霾形成机理 李克强：谁攻克，重奖谁！

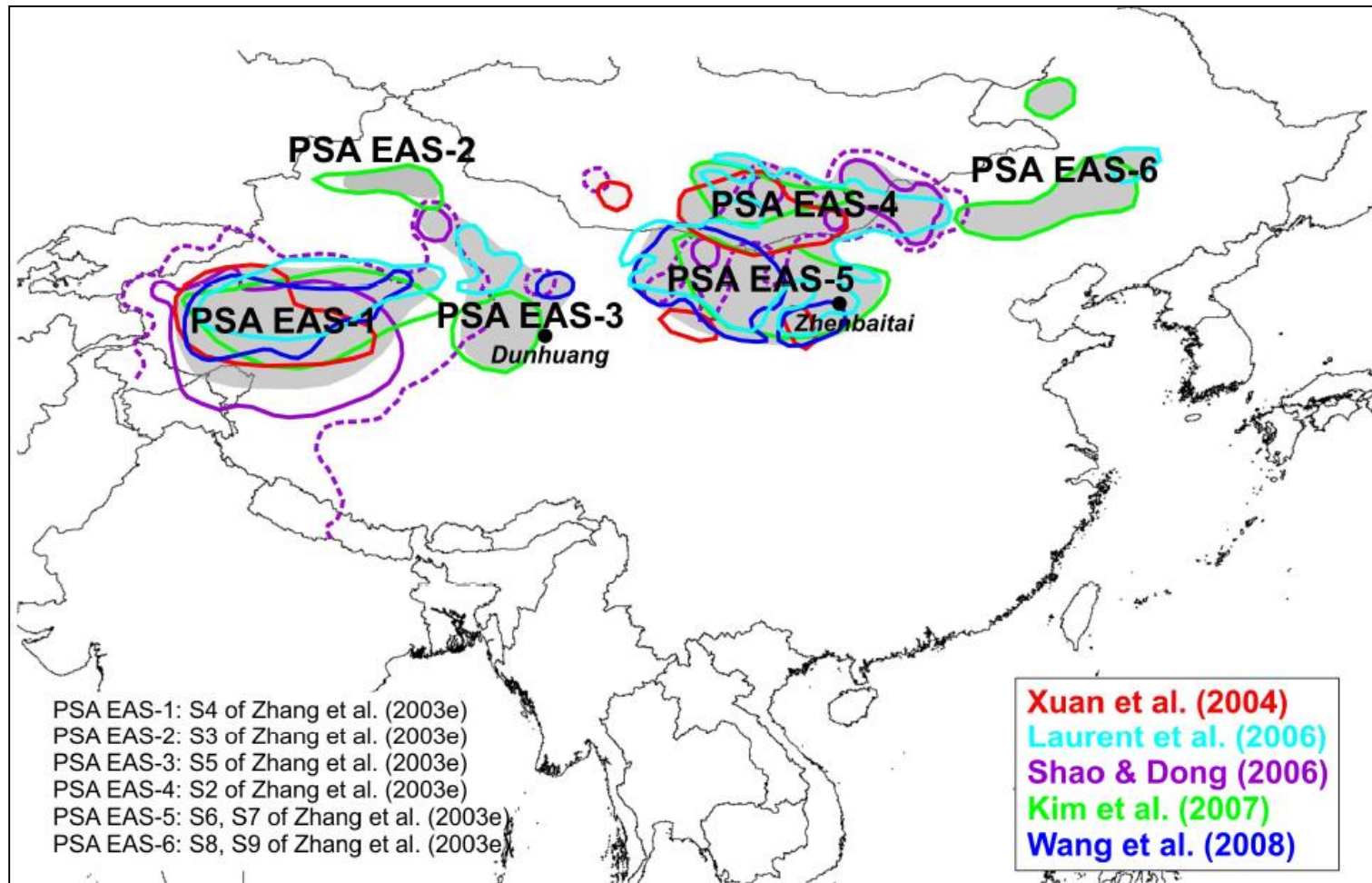
2017-03-10 | 【大中小】【打印】【关闭】



提 纲

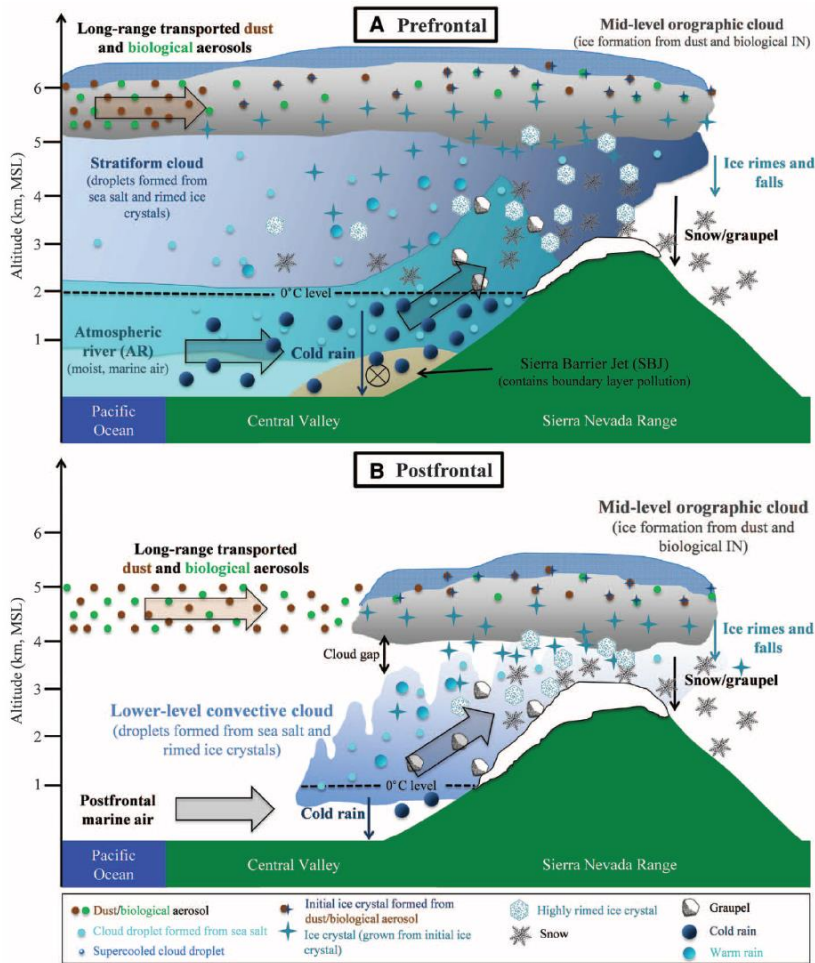
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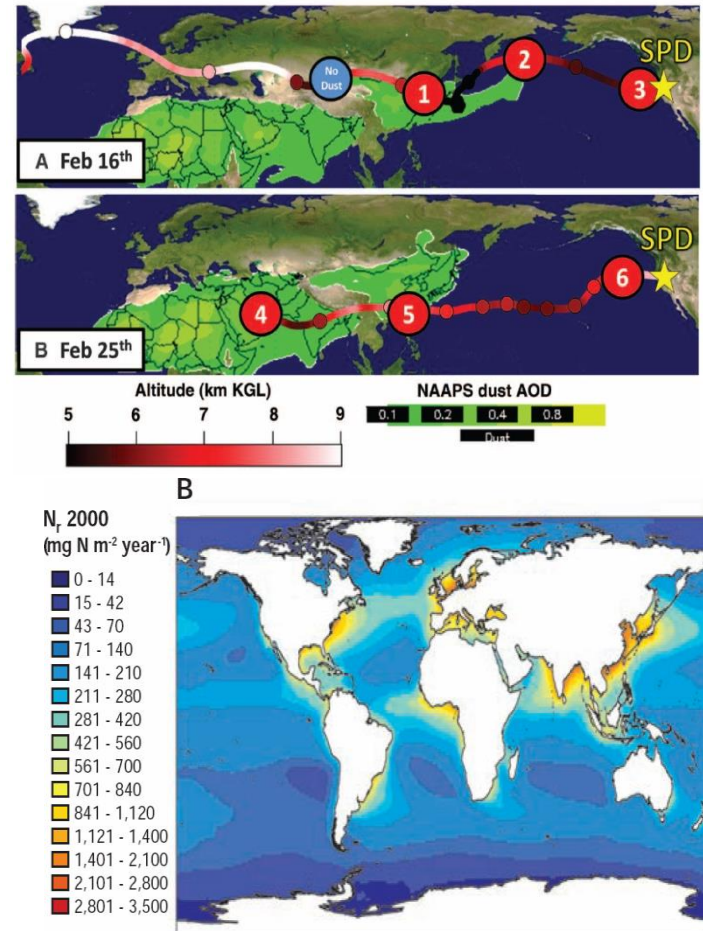


Taklamakan and Gobi desert areas are the major dust source regions of east Asia, releasing more than 800Tgyr^{-1}

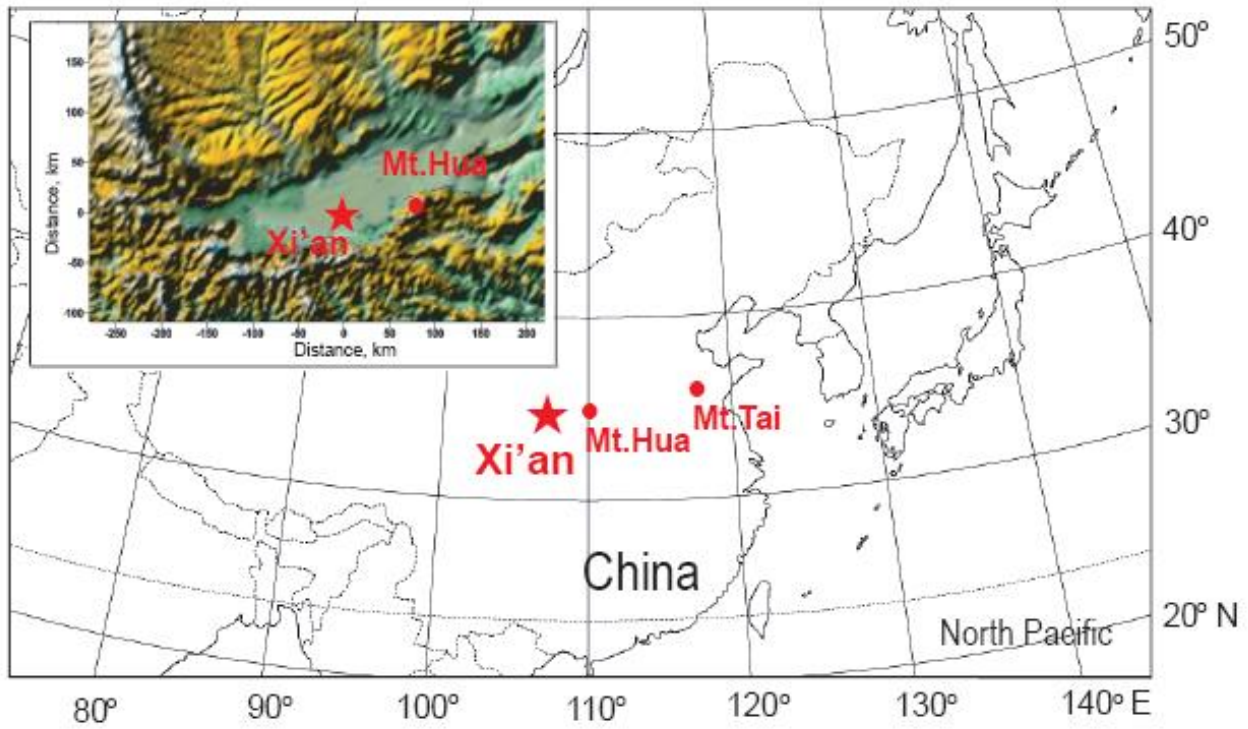
2. 沙尘暴期二次粒子



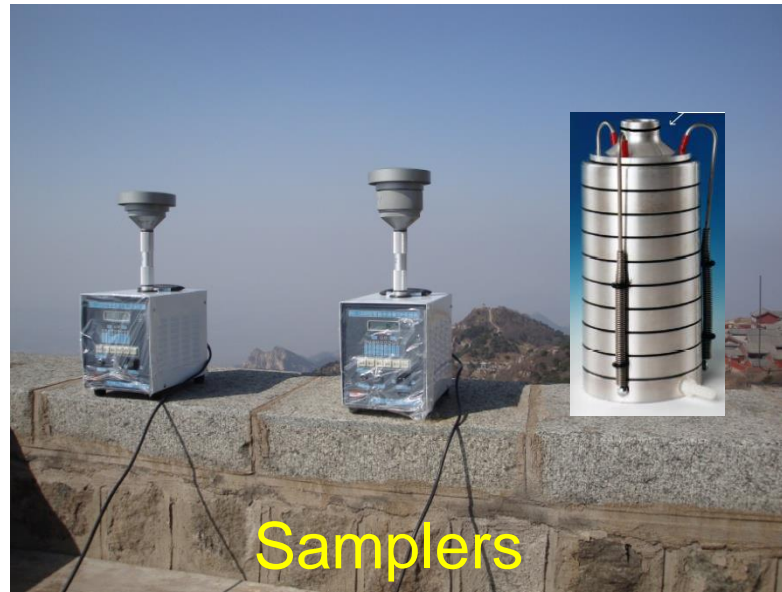
(Creamean et al. Science, 2013)



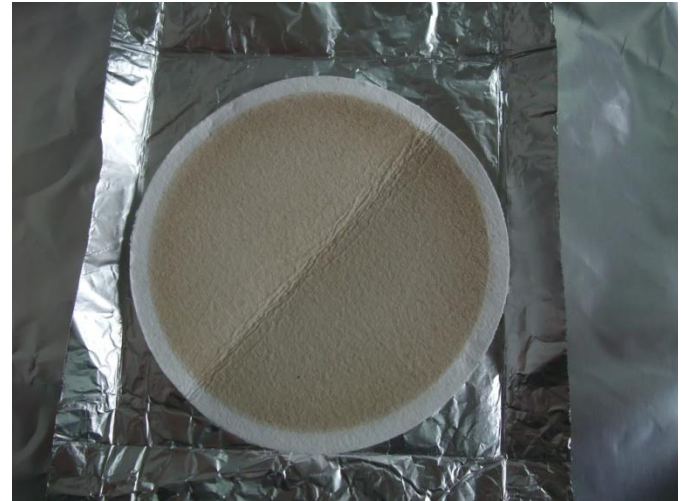
(Duce et al. Science, 2008)



Sampling sites



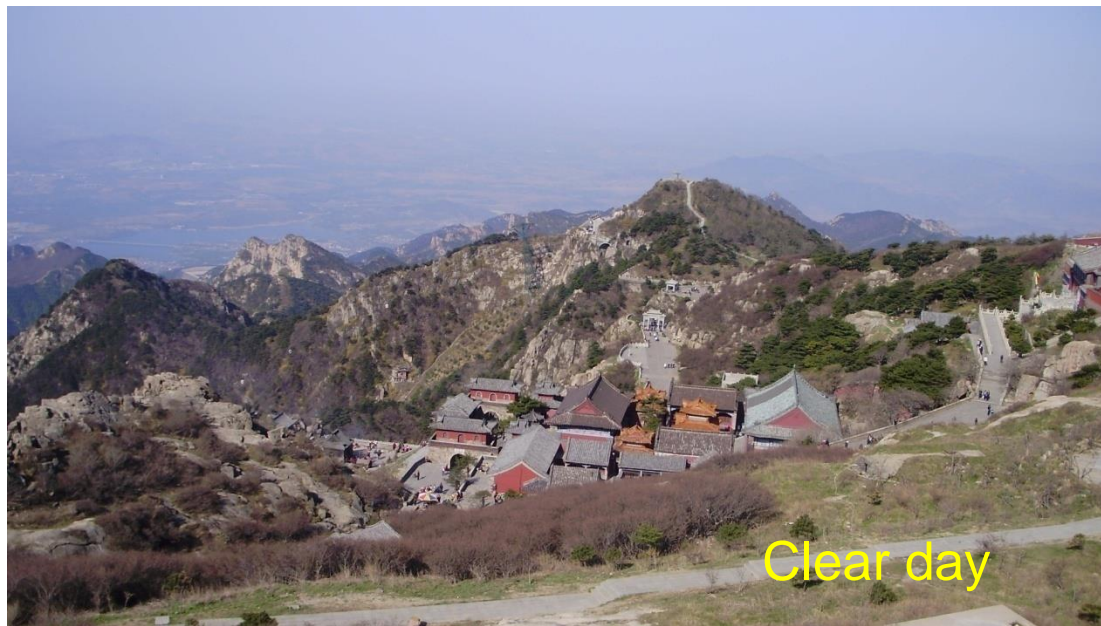
(i) Xi'an



(ii) Mt. Hua



(iii) Mt. Tai



I. Non-event days (Spring 2009)



Xi'an

Mt. Hua

Mt. Tai

II. Dust storm day (April 24, 2009)

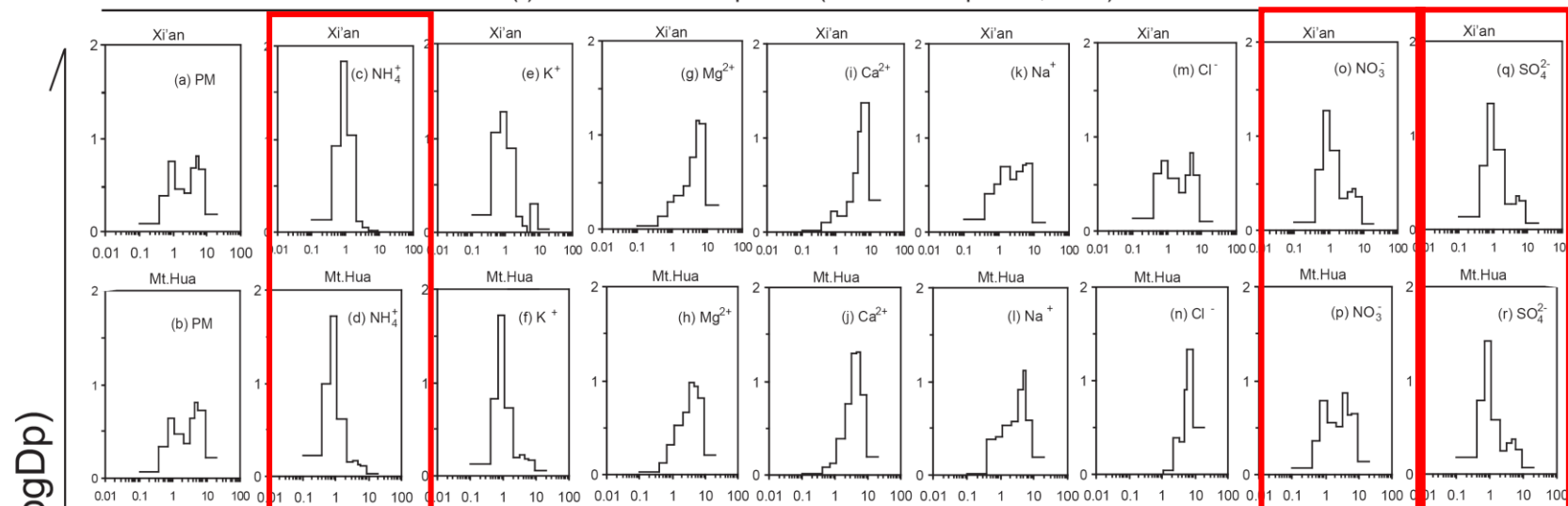


Xi'an

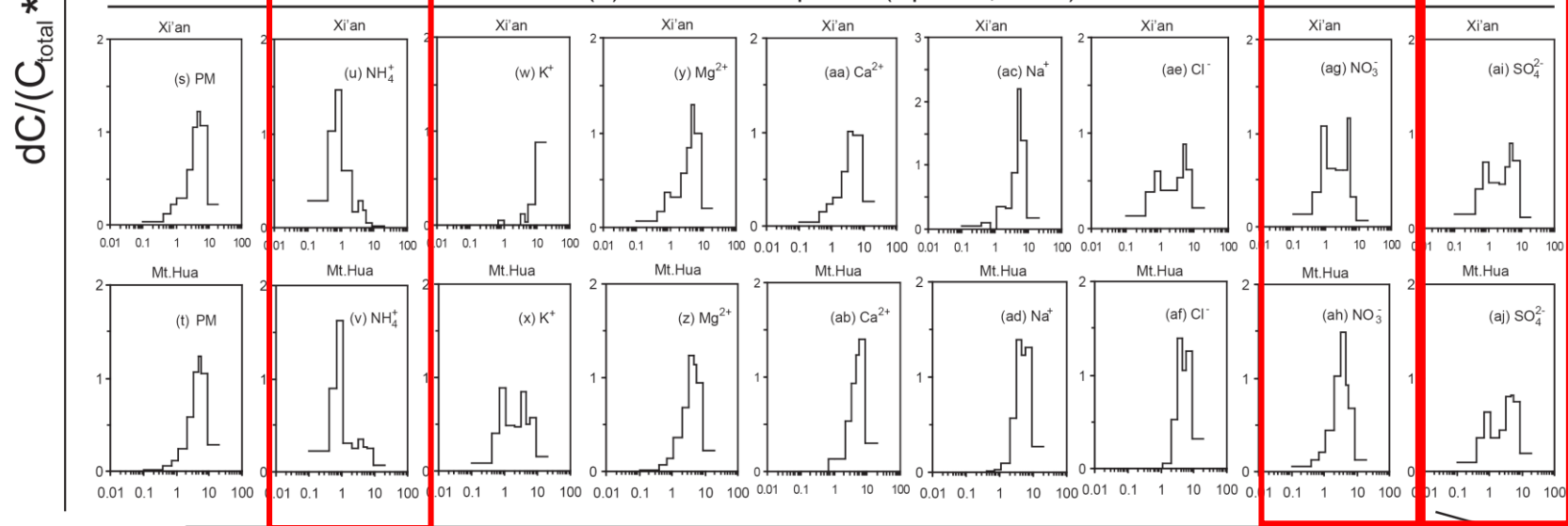
Mt. Hua

Mt. Tai

(I) Non-dust storm period (March 25-April 25, 2009)



(II) Dust storm II period (April 24, 2009)

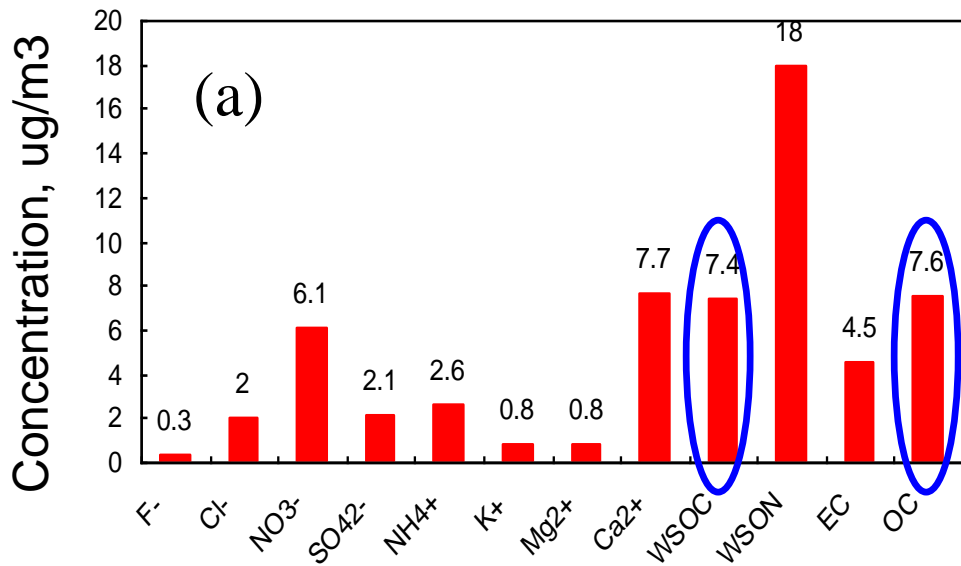
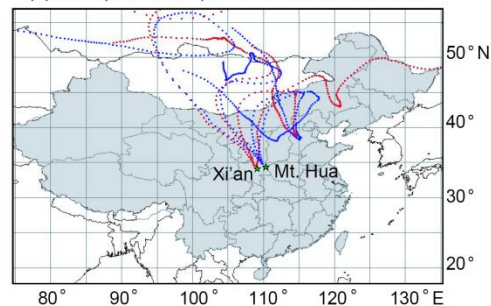


Particle size, D_p (μm)

(Wang* et al., Atmos .Chem. Phys. 2013)

Heterogeneous reactions in Xi'an during the DS II event

(c) DS II (04/24/2009) (Red: Xi'an; Blue: Mt. Hua)



Reaction rates, cm³molecules⁻¹s⁻¹

NO₂ + OH → HNO₃ k₁₇ = 8(-12)

SO₂ + OH → -- → H₂SO₄ k₁₈ = 6(-13)

"SO₂ + H₂O₂ + CLOUD → H₂SO₄" k₁₉ = 6(-16)

Rodhe & Crutzen, Tellus 33 (1981)

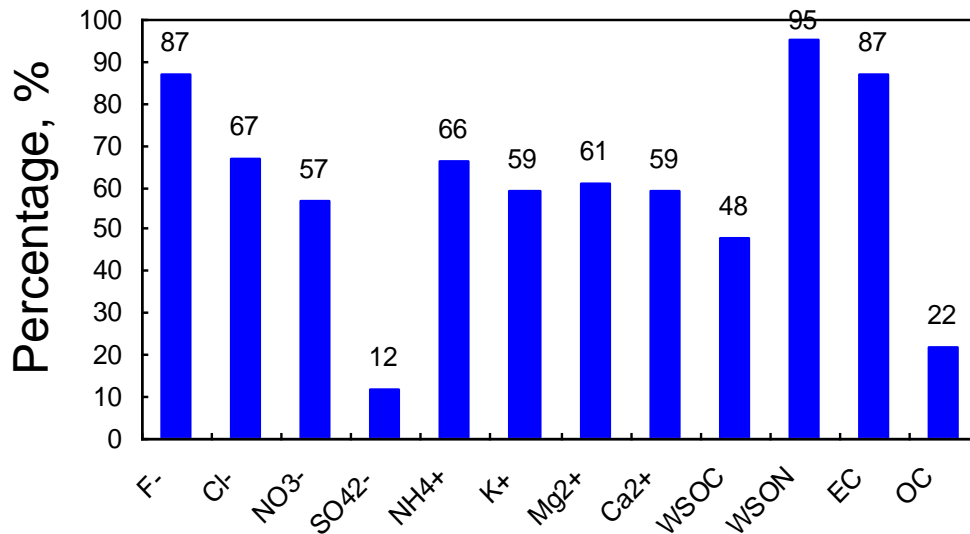
Contribution from local sources due to heterogeneous reactions

NO₃⁻ 6.1ug/m³, 57%

SO₄²⁻ 2.1ug/m³, 12%

WSOC 7.4 ug/m³, 48%

WSON 18 ug/m³, 95%



Contribution by long-range transport

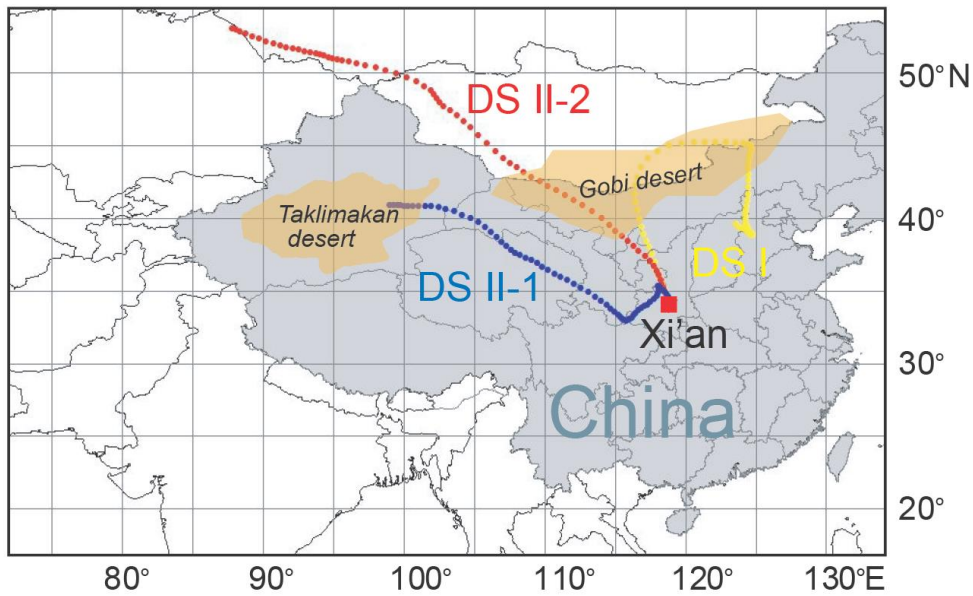
NO₃⁻ 4.6ug/m³, 43%

SO₄²⁻ 15.4ug/m³, 88%

WSOC 8.5 ug/m³, 52%

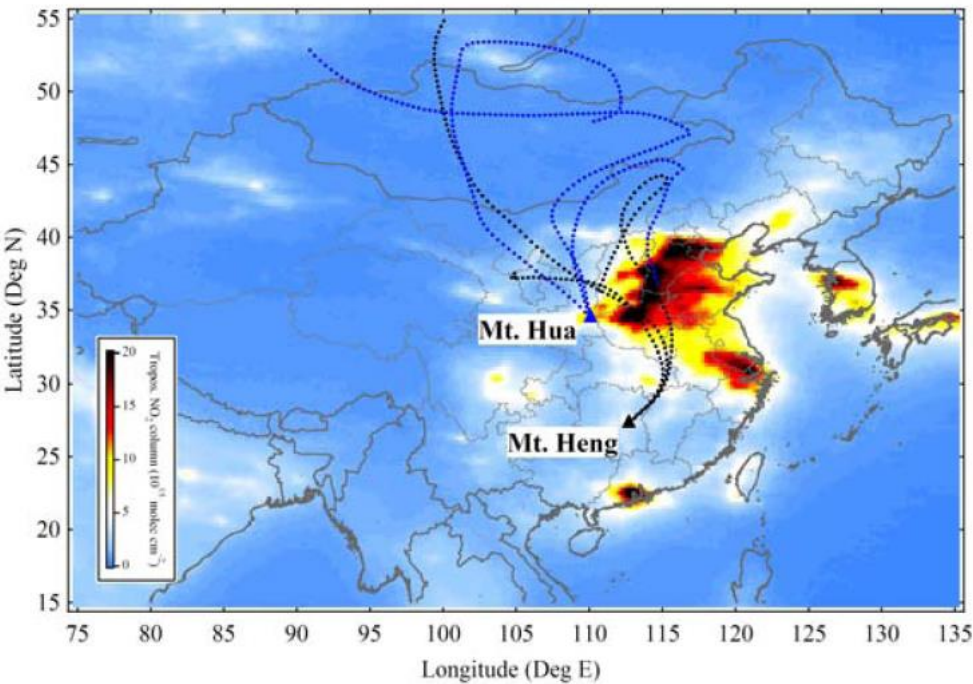
WSON 2 ug/m³, 5%

Fig. 8 Productions of pollutants from local sources in Xi'an



SP-equivalent concentrations^a of dicarboxylic acids, keto-carboxylic acids, ls, and other species during dust storm events in Xi'an, Central China

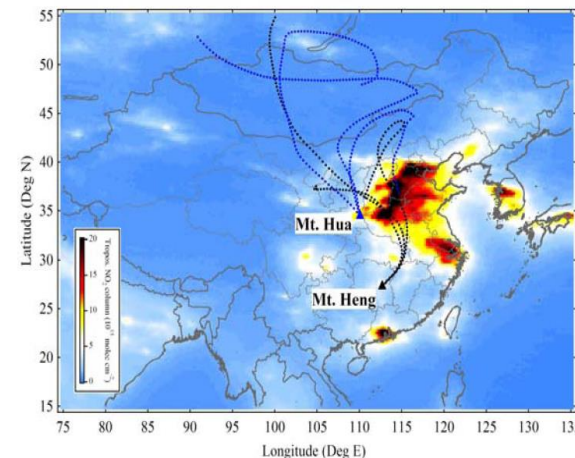
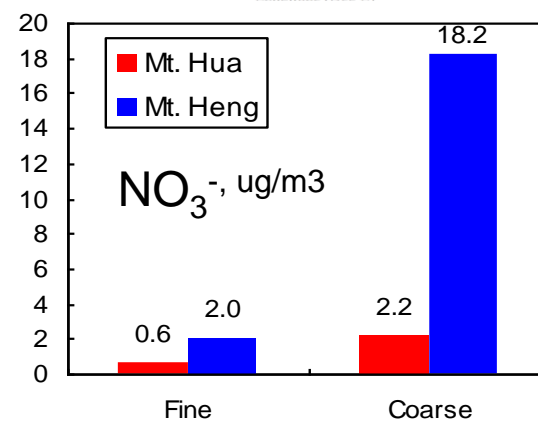
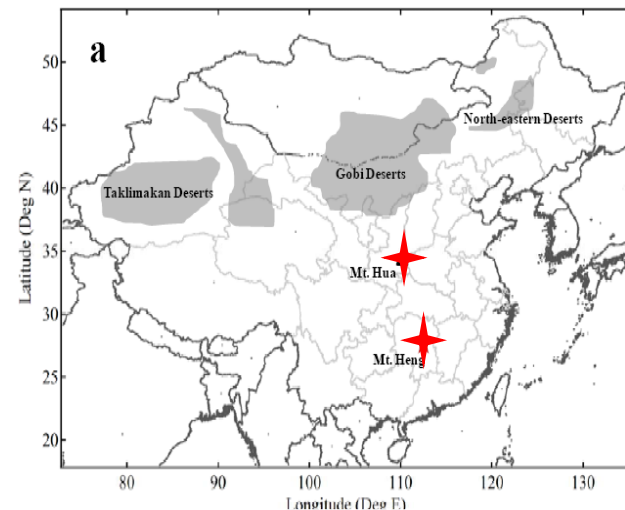
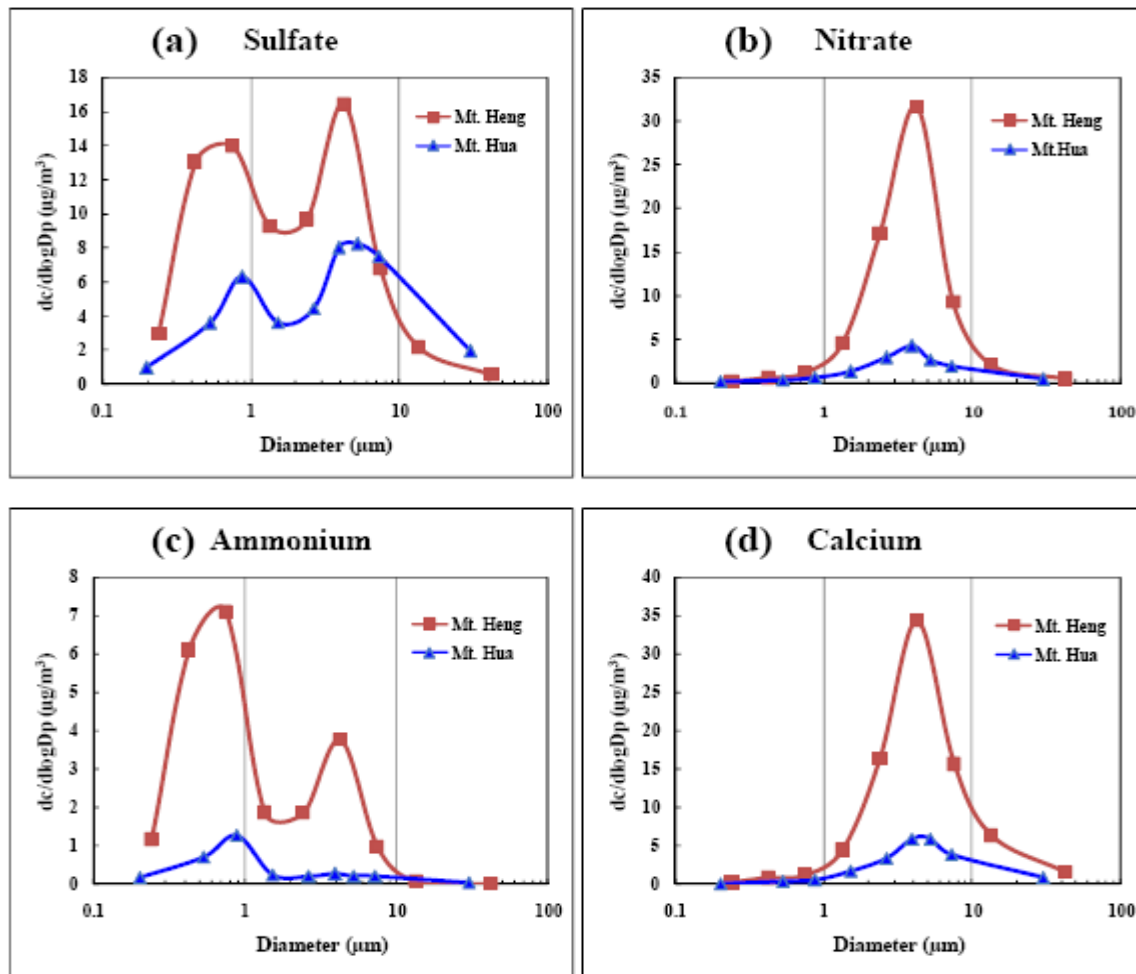
	DSI	DSII-1	DSII-2
I. Dicarboxylic acids, ng m ⁻³			
Oxalic (C ₂)	1511	539	365
Malonic (C ₃)	163	241	230
Succinic (C ₄)	200	117	93
Glutaric (C ₅)	50	19	19
IV. Inorganic ions, μg m ⁻³			
F ⁻	0.1	1	1
Cl ⁻	19	6	2
NO ₃ ⁻	22	9	4
SO ₄ ²⁻	32	28	17
NH ₄ ⁺	2	1	1
Na ⁺	6	26	32
K ⁺	3	1	1
Mg ²⁺	2	2	1
Ca ²⁺	11	27	9
Subtotal	97	100	67
V. Other species, μg m ⁻³			
WSOC ^b	38	46	38
WSIC ^b	27	28	13
Particle mass (PM)	682	1144	2406



(Wang et al., Atmos. Environ., 2015)

(Nie et al, Atmos. Chem. Phys. 2012)

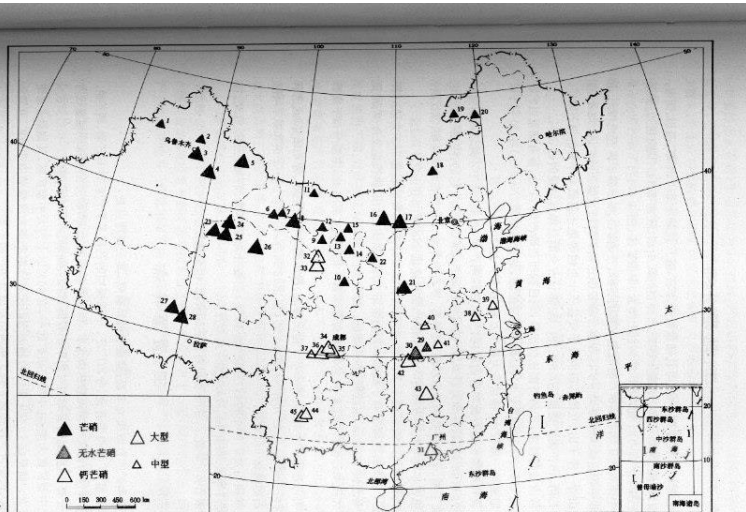
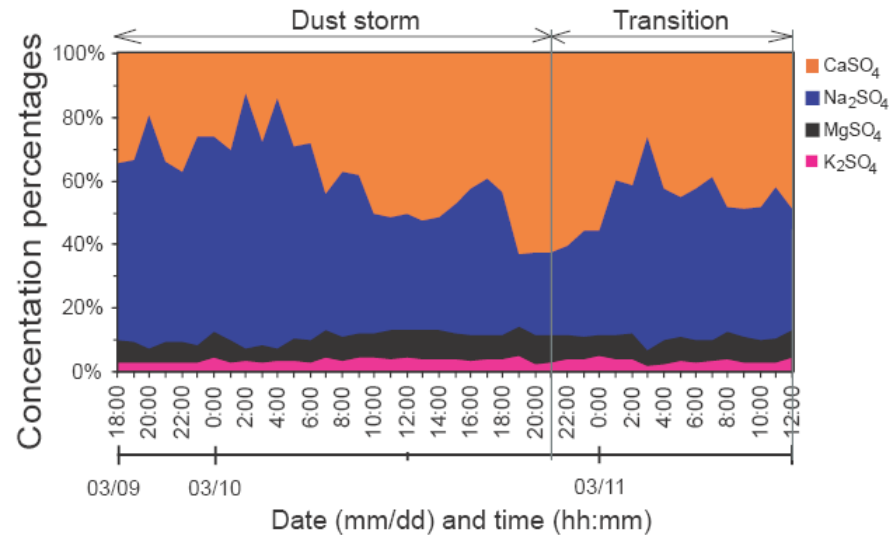
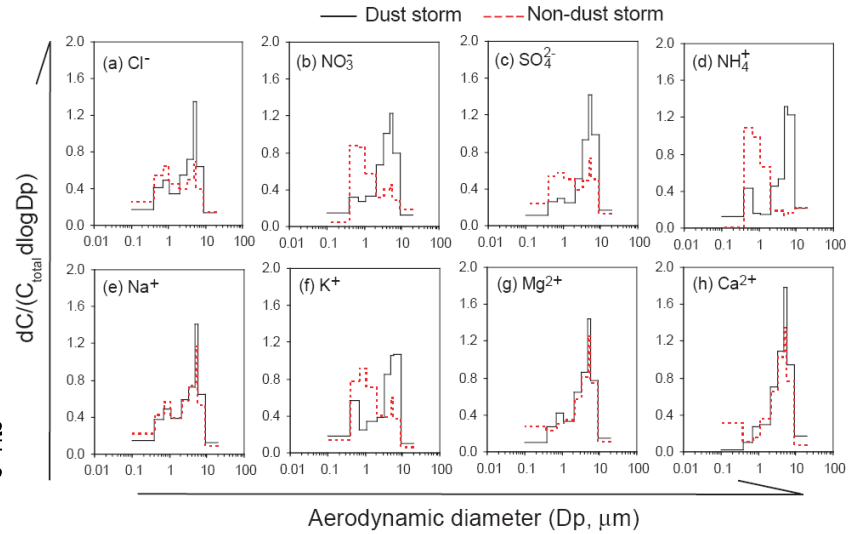
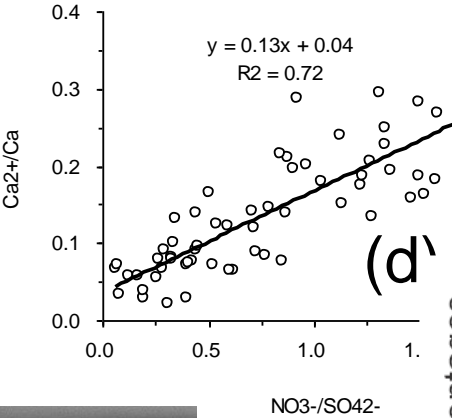
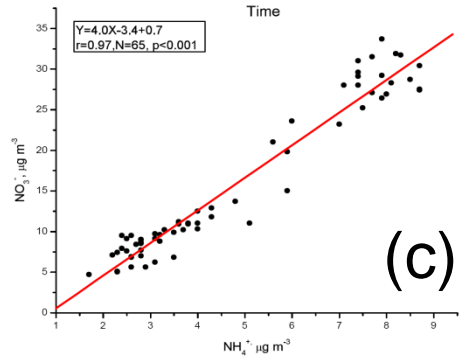
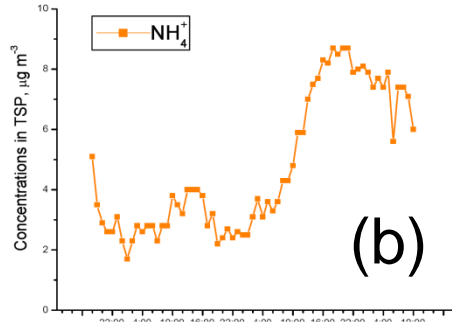
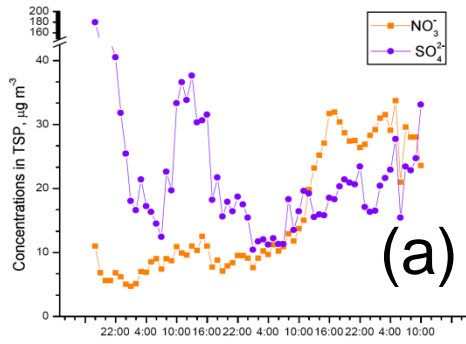
Enhanced heterogeneous formation of NO_3^-



$\text{PM}_{>1}$ $\frac{\text{NO}_3^-}{\text{SO}_4^{2-}}$ Mt. Hua 0.34
 Mt. Heng 1.4

(Nie et al., Atmos. Chem. Phys., 2012)

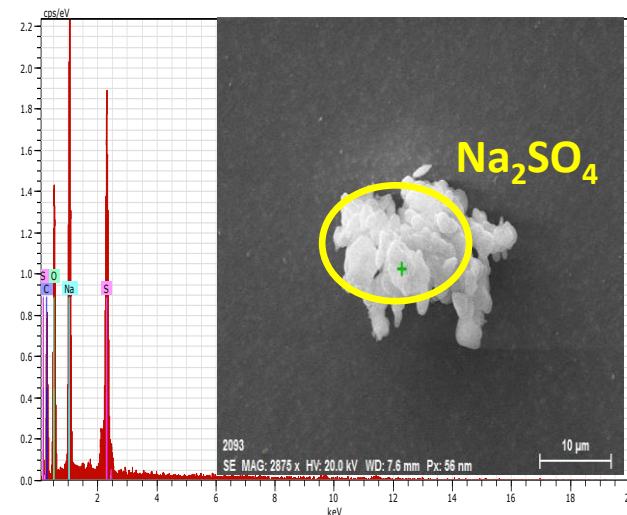
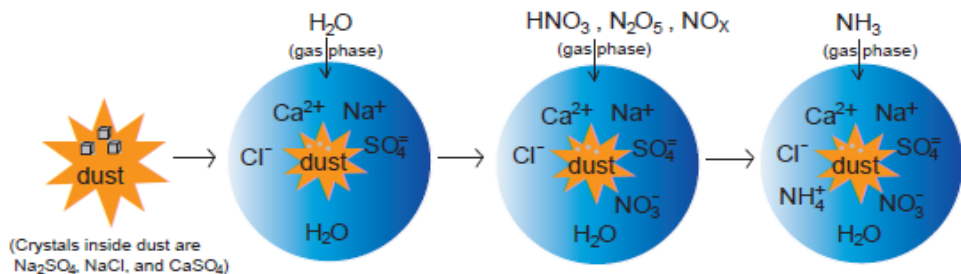
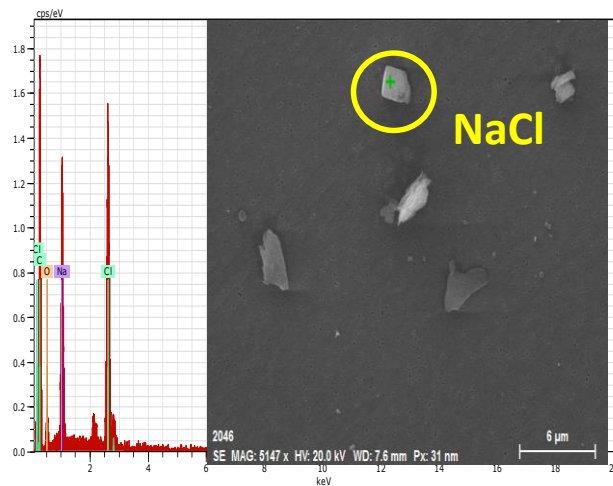
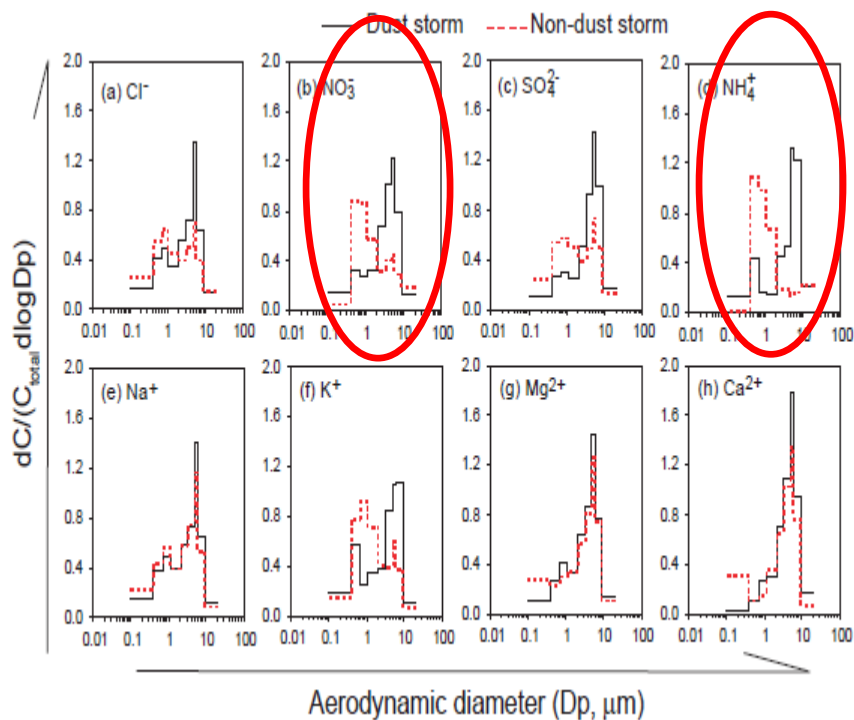
SO₄²⁻, NO₃⁻ and NH₄⁺ 小时浓度变化 (2013年西安沙尘暴期间, March 9-12, 2013)



(Wang et al. 2014, Atmos. Chem. Phys.)

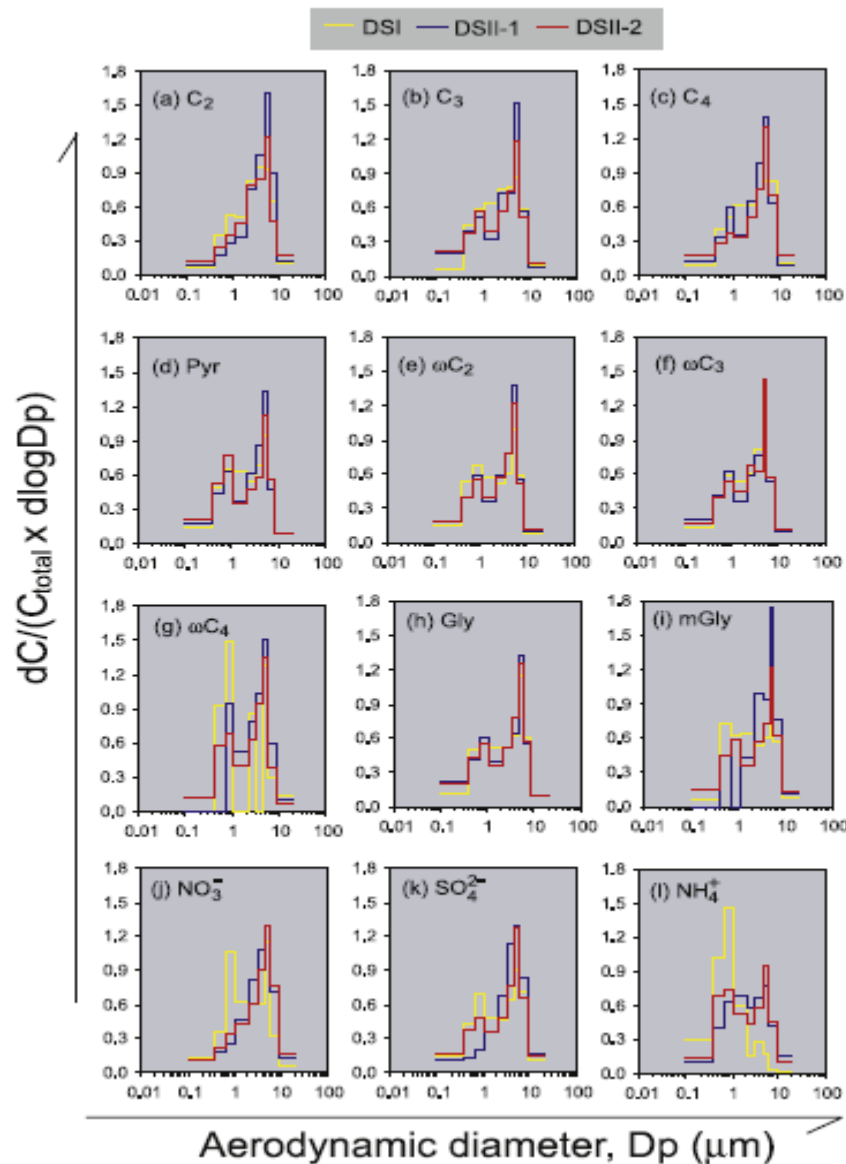
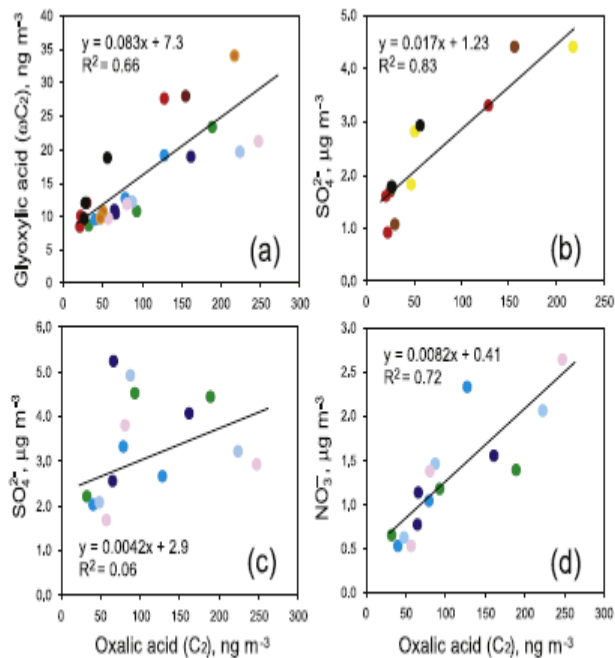
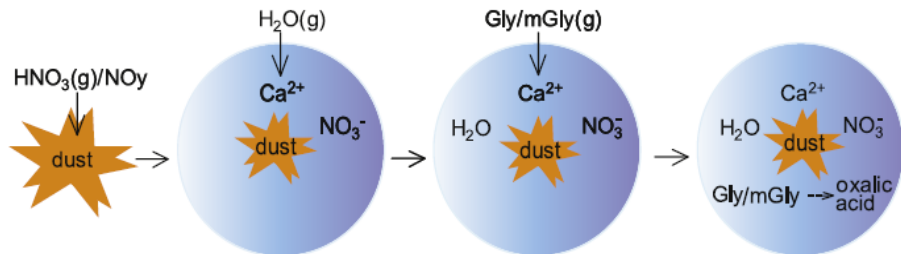
硝酸铵

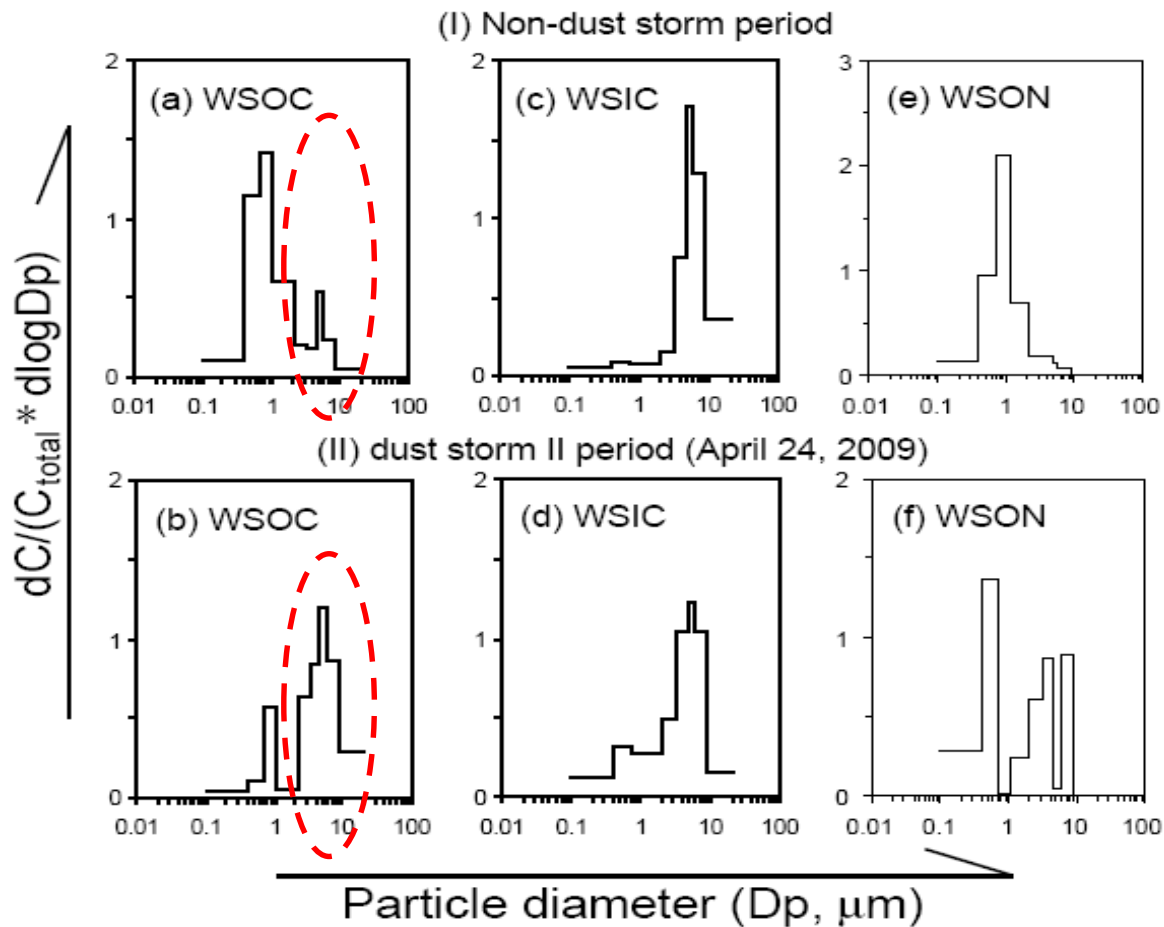
(Wang et al., ACP 2014, AE 2105,)
 (Zhang, Wang, et al., Chem. Rev. 2015)



沙尘表面二次有机气溶胶

(Wang et al., ACP 2014, 2017; AE 2105,)
(Zhang, Wang, et al., Chem. Rev. 2015)





	F^-	Cl^-	NO_3^-	SO_4^{2-}	NH_4^+	K^+	Mg^{2+}	Ca^{2+}	WSOC	EC	OC
Concentrations of species from local sources, $\mu\text{g m}^{-3}$	0.3	2.0	6.1	2.1	2.6	0.8	0.8	7.7	7.4	4.5	7.6
Relative abundance to the total in PM_{10} , %	87	67	57	12	66	59	61	59	48	87	22

(Wang* et al., Atmos .Chem. Phys. 2013)